



Effect of pesticide exposure on stunting incidence: a systematic review and meta-analysis

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As an endocrine disruptor chemical, pesticide exposure may affect the regulation of growth hormones such as insulin-like growth factor-1 (IGF-1). A few current studies to date have noted that long-term pesticide exposure disrupted IGF-1, a potential risk of stunting in children. This study aims to evaluate studies to date of the effect of pesticide exposure on stunting incidence. This systematic review and meta-analysis adhered to the PRISMA (Preferred Reporting Items for Systematic Review and Meta-Analysis) guidelines. The PubMed and EBSCO databases were searched for relevant articles without publication restrictions. This review aimed to include reviews, randomized controlled trials (RCT), and cohort, case-control, and cross-sectional studies, which provide actual exposure types of pesticides with stunting measurement by height-age *z* score. A screening, extraction, and synthesis were conducted, leading to a consensus for reaching mutual agreement. The analysis was performed using Microsoft Excel 2017 for the screening and extraction, Revman version 5.4 software for the meta-analysis, and OpenMEE software for the meta-regression. Of the 13 studies subjected to the qualitative analysis, 6 were eligible for inclusion in the meta-analysis: 2 reviews, 2 RCTs, 2 cohorts, 2 case-control, and 5 cross-sectional studies. Exposure to organophosphate and pyrethroid pesticides was not associated with stunting ($P=0.78$; odds ratio [OR], 1.11; 95% confidence interval [CI], 0.65–1.88). Heterogeneity existed for 79% of the meta-analysis ($P\leq 0.000$; $z=-5.37$; 95% CI, -0.034 to -0.016), and the meta-regression identified age as the causative covariate. Pesticide exposure, regardless of type, is not associated with stunting in children.

Key words: Pesticide exposure, Organophosphate, Pyrethroid, Stunting, Child

Key message

This review aimed to determine whether pesticide exposure is associated with stunting in children. The 13 included studies agree that pesticide exposure is not correlated with stunting incidence regardless of substance type (organophosphate and pyrethroid). Heterogeneity appeared with age covariate as potential confounding. The evidence of this study is challenging, as the adverse effects of pesticides grossly occurred. The protection of children is warranted for preventing future neurodevelopment issues.

Introduction

Over the past few decades, stunting has become a significant public health concern. The growth of many children is being stunted globally due to a complex interaction of household, environment, culture, and socioeconomic-political factors as reported by the World Health Organization's conceptual framework of childhood stunting.¹ Although the prevalence of stunting has decreased over the past 3 decades, an estimated 21.3% (144 million children) under 5 years of age are stunted globally.² The largest proportion of these children live in Africa, followed by a few parts of the Asian region and Latin America.² Like other causes of malnutrition in children, stunting affects a child's physical endurance and causes neurodevelopmental deficits and cognitive dysfunction.³ Moreover, a stunted child is prone to contracting infectious diseases⁴ and developing chronic diseases with long-term effects, such as obesity and hypertension.⁵

In addition to the aforementioned interactions, stunting may be influenced by the ingestion of food contaminated

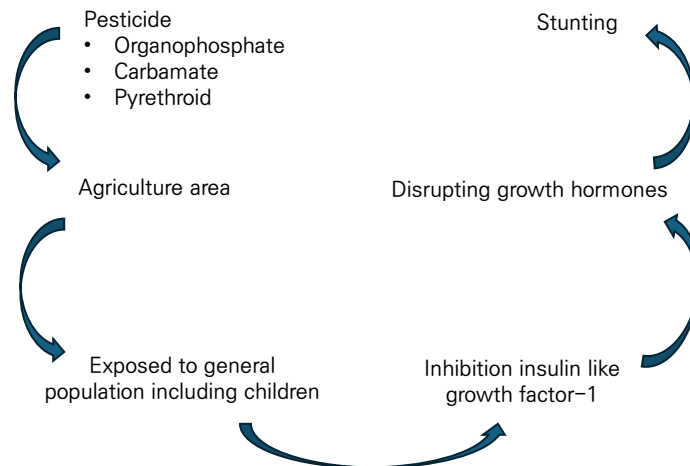
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Graphical abstract

by pesticides. An estimated 3 billion kilograms of pesticides are used across the globe annually,⁶ a figure that is expected to increase over the forthcoming year. Thus, the health effects of long-term exposure to pesticides require analysis. Long-term pesticide exposure plays a role in dysregulating hormone interactions, such as acetylcholinesterase⁷ (interrupting neurobehavioral disorders) and insulin-like growth factor-1 (IGF-1; disrupting growth hormone), in children.⁸ As a vulnerable population, children in the agriculture sector are at high risk of pesticide poisoning. The analysis of urine samples from children exposed to pesticides showed that levels of dimethyl thiophosphate were proportionately higher than those of other dialkyl phosphate metabolites.⁹ A similar study also reported that the mean urinary dimethyl organophosphate metabolite level was higher for children who lived <61 m from an orchard (0.1 µg/mL) than for those who lived >61 m away (0.04 µg/mL).¹⁰

The long-term effect of chemicals, particularly pesticide exposure, has been reported on behalf of stunted children.^{11,12} IGF-1 dysregulation is evidence of this mechanism through the inhibition binding-site receptors.¹³ Zumbado et al.¹⁴ reported IGF-1 concentrations in children and young adults exposed to organochlorine DDT pesticide. Among those participants, mean IGF-1 concentrations were higher in females than males (298 µg/L vs. 249 µg/L for young adults and 312 µg/L vs. 293 µg/L for children, respectively). Among the DDT metabolites, p,p'-dichlorodiphenyldichloroethylene (p,p'-DDE) was the higher than p,p'-dichlorodiphenyltrichloroethane (p,p'-DDT) and p,p'-dichlorodiphenyldichloroethane (p,p'-DDD) (80.6 ng/g-fat in boys and 60.40 ng/g-fat in young men). Thus, organochlorine metabolites, DDE, DDD, and total DDT were significantly associated with IGF-1 concentrations in young male, but not female, subjects. The authors concluded that organochlorine pesticides modulated IGF-1 differently by age and sex. Another survey by Kartin et al.¹² strengthened the previous

finding by Zumbado et al.¹⁴ that the mean IGF-1 level decreased twofold among children exposed to pesticides used for onion farming versus unexposed groups (66.74 ng/mL vs. 112.57 ng/mL, respectively).

Considering the impact of IGF-1 inhibition, pesticide contamination in children likely causes stunting. Among the pesticide classes, organochlorine pesticide toxicity has most often been reported for children, adolescents,¹⁴ and pregnant women.^{15,16} Conversely, despite limited data on the related mechanisms, exposure to the organophosphate and pyrethroid classes have also been associated with stunting.^{17,18} The massive use of the organophosphate class replaced organochlorine for agriculture and its applications are enormous; thus, studies examining the effect of exposure to organophosphate and other insecticides on children's development have become more common. A systematic review by Bliznashka et al.¹⁹ examined the effects of pesticide exposure on birth weight, birth length, low birth weight, and preterm birth. The study summarized that organochlorine and its metabolite were the most commonly used in agriculture-rich areas, followed by organophosphate, carbamate, and pyrethroid. The study found no association between prenatal exposure to pesticide, regardless of type, with outcomes including birth length.¹⁹ Thus, this systematic review and meta-analysis aimed to strengthen the previous evidence that exposure to pesticides, regardless of class, contribute to stunting among children. This study may provide essential insight and feedback to control pesticide-induced growth and development effects in children.

Methods

This systematic review and meta-analysis was based on PRISMA (Preferred Reporting Items for Systematic Review and Meta-Analysis) guidelines,²⁰ and its protocol number

Table 1. Predefined search terms

Databases: PubMed, EBSCO
(((((pesticide) OR (organophosphate) OR (organophosphorus) OR (insecticide) OR (herbicide) OR (fungicide) OR (profenofos) OR (chlorpyrifos) OR (organochlorine) AND (((stunting) OR (stunted growth) OR (malnutrition) OR (malnourished) OR (failure to thrive) OR (inadequate nutrition) OR (dwarf) OR (retarded growth))

was registered in PROSPERO (CRD42023462432).

1. Study selection

Eligible studies published in English through April 2023 were retrieved from the PubMed and EBSCO databases. The reference lists of the retrieved articles were manually searched for additional applicable sources. We used predefined search terms to identify relevant articles (Table 1). After the removal of duplicates, the titles and abstracts were screened by the authors (MF, MK, VW, KW, NAD, MR, and ESH). Articles that met the inclusion criteria were then subjected to full-text screening. Cases of discrepancy were settled by consulting other authors (SRS and YAJ). Consensus was reached upon mutual agreement of all authors. A paper's corresponding author was contacted for clarification if needed. The data collection and extraction processes were conducted using Microsoft Excel 2017 and Mendeley software, and the data were stored in the cloud using a service provided by Universitas Islam Indonesia.

2. Inclusion and exclusion criteria

The inclusion criteria were the analysis of the influence of pesticide exposure, regardless of class, on stunting incidence. Pesticide exposure was defined for every child who lived on or near an agricultural facility. We broadened the exposure criteria based on laboratory examinations and observational criteria to capture exposed and nonexposed children. Additionally, stunting in an article must be quantified by a height-age *z* score (HAZ) measurement, while self-reported information or other measurements were omitted from the analysis. Observational studies (cohort, case-control, and cross-sectional), randomized controlled trials (RCTs), interventional studies, reviews, and proceedings were eligible for inclusion and analysis. However, opinion, perspective, and descriptive data, including case reports, case series, and case studies, that did not examine the association between pesticide exposure and stunting were excluded. A certainty analysis of every included study was performed using the Grading of Recommendations Assessment, Development, and Evaluation (GRADE) summary.²¹⁾

3. Quality assessment

A quality assessment was performed using the Newcastle-Ottawa Scale (NOS) for non-RCTs, observational studies,

cohort, and case-control studies.²²⁾ This method offers a risk of bias (RoB) assessment in 3 domains: selection, comparability, and outcome or exposure. Given the limitation of poor agreement and lack of a comprehensive manual,²³⁾ the NOS method is the most applicable RoB tool for nonrandomized studies and meta-analyses. Cross-sectional studies were examined using the Agency for Healthcare Research and Quality guidelines.²⁴⁾ RCT studies were examined using the Cochrane risk-of-bias tool.²⁵⁾ Two authors (YAJ and MK) performed the RoB analysis.

4. Extraction and synthesis

Seven authors (MF, MK, VW, KW, NAD, MR, and ESH) extracted each study's author name, research design, population, pesticide type, stunting outcome, and finding. In cases of uncertainty, other authors were consulted (SRS and YAJ) until consensus was reached. The data were analyzed similarly. The data analysis focused on the endpoint estimation effects, including relative risk (RR), odds ratio (OR), mean difference, correlation coefficient, *P* value, and confidence interval (CI). Studies for which statistical data were missing were removed from the quantitative analysis. A study's corresponding author was contacted as necessary for further information. If no OR or RR values were reported by a study, the calculation of a 2x2 table was warranted. Since this study analysis was also quantitative, heterogeneity was used to assess the diversity of the included studies. A random effects model was used, and a forest plot was created to display effect estimates and CIs for the individual studies. The source of any heterogeneity was identified by meta-regression using OpenMEE software. A sensitivity and validity analysis was conducted using a funnel plot in Revman version 5.4 software (Cochrane, Fredericksburg, VA, USA).

Results

The database search retrieved 3,178 papers related to the effects of pesticide exposure on stunting incidence. After the elimination of duplicates, 2,753 papers remained. The elimination of studies with irrelevant topics resulted in 20 eligible papers. Of them, 13 were included in the systematic review and meta-analysis. A detailed flowchart of the study retrieval and selection process is presented in Fig. 1.

1. Baseline characteristics of included studies

A total of 13 studies were included in the analysis: 2 reviews,^{26,27)} 2 RCT,^{17,28)} 2 cohort studies,^{18,29)} 2 case-control studies,^{12,30)} and 5 cross-sectional studies.^{11,31-34)} Four studies elucidated the effects of pesticides on stunting,^{11,12,27,30)} 3 focused on organophosphate and carbamate,^{29,31,32)} 5 focused

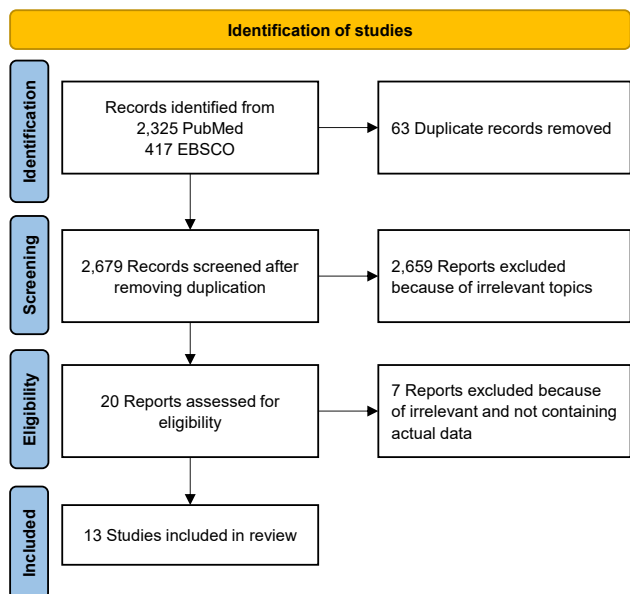


Fig. 1. Flowchart diagram retrieved papers.

on permethrin, a pyrethroid class,^{17,18,33,34} and one simultaneously analyzed metabolite herbicide, organophosphate, and pyrethroid.²⁹ Most studies analyzed stunting using HAZ; only one used the ponderal index³⁵ and, thus, was omitted from the further analysis. Among the 13 retrieved studies, 6 included a meta-analysis. The detailed baseline characteristics of the included studies are presented in Table 2 and Supplementary Table 1.

2. Effects of pesticide exposure on stunting

Kartini et al.¹² examined high-exposure pesticide-induced stunting among school-aged children who lived near onion farming (adjusted OR [aOR], 3.9; 95% CI, 1.15–13.26). Evidence of stunting was proven by HAZ and low IGF-1 levels (aOR, 8.35; 95% CI, 3.65–19.14). Kartini et al.¹² thoroughly evaluated risk factors for stunting, including nutritional status, low birth weight, infection history, and parents' education and occupation. Protein and energy

Table 2. Baseline characteristics of retrieved studies

Study	Study design	Participants	Exposure	Outcome			Summary
				Stunting	Total sample	Estimation effect	
Salam et al. ²⁶⁾	Scoping review	1 Study included (411 children with ITN and 420 control)	ITN (permethrin)	NI	NI	OR, 1.11; 95% CI, 0.86–1.42	There was no association between ITN used with stunting
Vilcin et al. ²⁷⁾	Systematic review	1 Study included consist of 354 children	Pesticide	NI	NI	NI	Pesticide was associated with stunting
Friedman et al. ²⁸⁾	RCT	455 Children in ITN village and control 421 children.	ITN (permethrin)	155	876	Mean, -0.96; 95% CI, -1.0 to -0.86	Pesticide was not associated with stunting
Ter Kuile et al. ^{17), a)}	RCT	411 Children in ITN intervention and 420 control	ITN (permethrin)	NI	831	MD, 0.36; 95% CI, 0.10–0.62; <i>P</i> =0.01	ITN was associated with stunting 0 to 23 months old children
Jaacks et al. ^{29), b)}	Cohort	289 Pregnant women followed for one-year exposed pesticide on stunting in her children	Herbicide, organophosphate, pyrethroid	197	289	RR, 0.99, 95% CI, 0.77–1.27 (TCPy)	Metabolite organophosphate, TCPy, did not correlate with stunting during 1-year follow-up
Arinaitwe et al. ¹⁸⁾	Cohort	99 HIV-unexposed 202 HIV-exposed 57 HIV-infected	ITN (permethrin) among children with HIV	252	358	RR, 1.24; 95% CI, 1.03–1.48; <i>P</i> =0.008	ITN exposure was associated with moderate severe stunting
Kartini et al. ¹²⁾	Case-control	160 Children 8-12 yr (48 case and 112 control)	Pesticide	48	160	OR, 3.9, 95% CI, 1.15–13.26	History high-exposed pesticide associated with stunting
Paudel et al. ³⁰⁾	Case-control	236 Children 6 to 59 months	Pesticide	118	236	aOR, 3.51, 95% CI, 1.33–9.23	Pesticide exposure was associated with stunting.
Grandjean et al. ¹¹⁾	Cross-sectional	72 Children primary school	Pesticide	31	72	<i>B</i> =-0.74, <i>SE</i> =0.35, <i>P</i> =0.03	Stunting was independent predictor of neurobehavior in exposure pesticide.
Friedman et al. ³³⁾	Cross-sectional	1862 Children	ITN (permethrin)	466	1396	Mean=-0.29, <i>P</i> =0.003	ITN used in high density parasitemia was associated with stunting
Amoako Johnson ³⁴⁾	Cross sectional	NI (using DHS data)	ITN (permethrin)	NI	NI	OR, 0.192; 95% CI, 0.18–0.199 (pooled data)	ITN coverage was associated with stunting in DHS data
Orozco et al. ³¹⁾	Cross sectional	54 = more intensive 40 = intermediate 41 = less intensive	Organophosphate and carbamate	49	135	<i>P</i> =0.019	There was association between intensity used pesticide with stunting in children
Handal et al. ³²⁾	Cross sectional	283 children aged 3-61 mo	OP and carbamate	151	283	OR, 0.04; <i>P</i> =0.87	Pesticide exposure did not correlate with stunting

DHS, demographic health survey; ITN, insecticide treatment net; NI, no information; MD, mean difference; TCPy, 3,5,6-trichloro-2-pyridinol; OR, odds ratio; aOR, adjusted OR; RR, relative risk; CI, confidence interval.

^{a)}Data were provided in mean difference of height-age z score among children 0–23 months of age during ITN intervention. ^{b)}Data were used for quartile 2 for 1-year follow-up of 3,5,6-trichloro-2-pyridinol exposure.

intake as well as high exposure to pesticides were associated with stunting. However, the sample size was inadequate, which limited the study's power. The study of Paudel et al.³⁰ corroborated the findings of Kartini et al.¹² in their adjusted multivariate analysis (aOR, 3.51; 95% CI, 1.33–9.23). The strength association of the results was proven by its adequate sample size, but the study was limited by its use of poor methods of detecting pesticide exposure. A review by Vilcins et al.³⁰ summarized a similar result to that of Paudel et al.²⁷ of pesticide-induced stunting (aOR, 3.51; 95% CI, 1.33–9.23). Apart from the previous findings, Grandjean et al.¹¹ found that a low score on the Stanford-Binet Copying Test, a neurodevelopment test, was associated with pesticide exposure ($\beta=-1.01$; standard error [SE]=0.37; $P<0.00$) and stunting in children ($\beta=-0.74$; SE=0.35; $P<0.05$). Of 37 exposed groups (-1.84 ± 0.83 ; range, -3.65 to 0.35), 15 stunted children had median dimethyl metabolite and diethyl metabolite levels of 0.2 nmol/kg/day and 0.27 nmol/kg/day, respectively.¹¹

In general, pesticide exposure was associated with stunting. However, the limited sample sizes and inadequate pesticide exposure assessments prevented acceptance of the stated hypothesis.

3. Effects of organophosphate exposure on stunting

Low use of organophosphate and carbamate was correlated with HAZ in children compared to their high and intermediate use (-1.95 ± 1.86 , -1.05 ± 1.5 , -1.25 ± 1.27 , $F=4.06$, $P=0.0194$, respectively).³¹ This finding seems different because the low carbamate and organophosphate use levels were 0.65 ± 0.93 kg and 0.64 ± 0.74 kg, respectively, compared with the high (0.28 ± 0.27 and 0.60 ± 0.47) and intermediate (0.45 ± 0.33 and 0.91 ± 0.79) levels.³¹ Additionally, significantly more carbamate was used than organophosphate ($P=0.04$ vs. $P=0.12$). Overall, stunting was noted in 11 of 41 children in the low use group versus 2 of 40 in the intermediate use group and 4 of 54 children in the high use group. Unfortunately, this study did not examine the correlation between urinary pesticide exposure and HAZ in children.³¹ Handal et al.³² also reported that organophosphate and carbamate pesticide exposure was not correlated with stunting incidence in children (OR, 0.04; $P=0.84$). Stunting, however, affects a child's gross motor function, making it a moderate variable or independent factor in an individual's neurodevelopment. The prevalence of stunting was increased after high or low pesticide exposure, and children in both regions replace with in both exposure level resided in rural areas with lower education and income levels.³² Another study reported on the effects of exposure to 4 organophosphate metabolites on stunting,²⁹ including 3,5,6-trichloro-2-pyridinol (TCPy; chlorpyrifos metabolite), 4-nitrophenol (parathion and methyl parathion metabolite), malathion dicarboxylic acid

(malathion metabolite), and 2-isopropyl-4-methyl-6-hydroxypyrimidine (IMPY; diazinon metabolite). This study found that organophosphate exposure was not correlated with stunting, even after 2 years of follow-up (RR at Q4: TCPy, 1.16 [0.8–1.68]; 4-nitrophenol, 0.85 [0.61–1.19]; IMPY, 1.04 [0.73–1.49]).²⁹ Exposure to organophosphate, regardless of class, was not associated with stunting. However, this study was limited by its small sample size, exposure assessments used, and poor design. Jaacks et al.²⁹ used the most applicable research method that seemingly required a larger sample size and long-term cohort evaluation to prove the relationship between exposure and outcomes.

4. Effects of pyrethroid exposure on stunting

Permethrin, a class of pyrethroid, was the most commonly used pesticide for treating malaria, namely insecticide-treated nets (ITN), according to some retrieved studies.^{17,18,29,33,34} Stunting cases increased parallelly with high exposure to malaria among children with HIV (RR, 1.24 [1.06–1.48]),¹⁸ malaria treatment with ITN caused malnutrition, including stunting (-0.20 ± 0.09),³³ and ITN coverage was less effective at preventing stunting (OR, 0.02 or mean difference, 0.19; range, 0.18–0.19).³⁴ An RCT by Friedman et al.²⁸ found that use of the ITN intervention did not correlate with stunting (mean pooled data, -0.95 ; 95% CI, -1.07 to -0.86). A similar RCT by Ter Kuile et al.¹⁷ indicated that use of the ITN intervention versus control was correlated with HAZ (mean ITN, -1.40 vs. control, -1.76 ; mean difference, 0.36; range, 0.10–0.62; $P=0.001$). Another study of the more specific metabolite pyrethroid 3-PBA found no correlation with stunting incidence at the first or second follow-up (mean difference, 1.01; range, 0.83–1.24; and mean difference, 0.95; range, 0.69–1.32, respectively).²⁹

Although an RCT design and large sample size were proposed in the ITN study on stunting, the result regarding the relationship between exposure and outcome was debatable. No suitable pyrethroid assessment was performed in either of the studies, so a questionnaire was used to collect the information. The robust method used in the Jaacks et al. cohort study inadequately proved an association between pesticide contamination and stunting. Thus, pyrethroid exposure was not associated with stunting incidence.

5. Meta-analysis

Six studies were included in the quantitative analysis, as presented in Figs. 2 and 3.^{11,12,28,30–32} Of 445 subjects who used ITN treatment to eradicate malaria in the study of Friedman et al.,²⁸ 79 (16.7%) denoted stunting, and 76 of 393 controls were positive for stunting. The study of Grandjean et al.¹¹ reported that 15 of 37 children with prenatal pesticide exposure versus 16 of 35 unexposed children were stunted. A history of pesticide applications was stratified

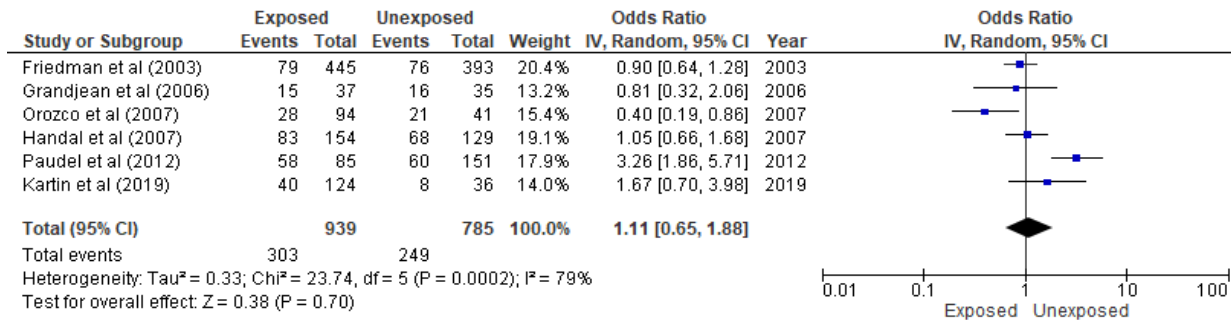


Fig. 2. Forest plot of association between pesticide exposure with stunting. IV, inverse variance; CI, confidence interval; *df*, degrees of freedom.

as high, moderate, low, and unexposed; of them, 13, 14, 13, and 8 were stunted, respectively. On the other hand, similar results were seen among control subjects (14, 31, 39, and 28, respectively). Thus, 40 of 124 exposed versus eight of 36 controls were stunted.¹²⁾ Orozco et al.³¹⁾ stratified pesticide use intensity into more, intermediate, and less levels. We defined high pesticide exposure as high and moderate intensity and low exposure as less intensity. The authors reported that 28 of 94 children with high exposure versus 21 of 41 children with low exposure were stunted.³¹⁾ In a flower plantation area in the study of Handal et al.,³²⁾ greater pesticide exposure led to stunting in 83 of 154 children, while lesser exposure led to stunting in 68 of 129 children. Paudel et al.³⁰⁾ also clarified the presence of 58 of 85 stunted children in pesticide-exposed groups versus 60 of 151 stunted children in nonexposed groups.

Total events were 278 of 845 population in the exposed group versus 228 of 774 in the unexposed group. Although the funnel plot to verify sensitivity confirmed symmetrical data, heterogeneity existed ($I^2=75\%$, $P=0.003$). Additionally, a meta-regression determined that age was a significant ($P\leq 0.001$) source of heterogeneity. Therefore, pesticide exposure was not associated with stunting in children ($P=0.28$).

Discussion

General exposure to pesticides was associated with stunting incidence. There are several points related to this result. First, the exposure assessment was more community-based than the individual assessment, and the community data could not provide actual concentrations compared to the individual data. Second, regarding baseline characteristics among the retrieved papers, particularly occupational status, most respondents worked as merchants¹²⁾ and housewives,²⁷⁾ which likely creates a risk of selection bias. Assessments of community exposure should focus on farmers, who work more frequently in the field. Second, the data were collected by interview or general assessment exposure^{12,27,30)} and included limited sample sizes,¹¹⁾ which likely created selection bias.

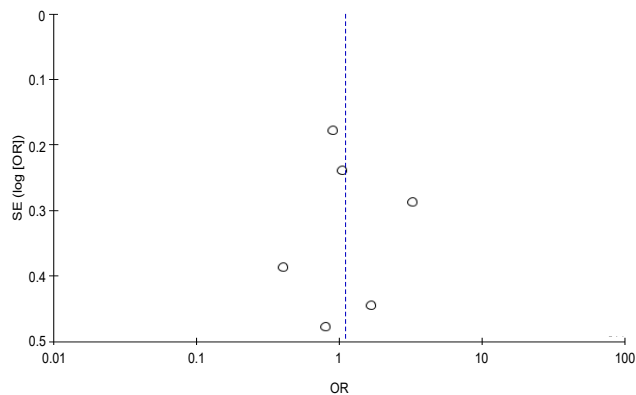


Fig. 3. Funnel plot of retrieved studies. SE, standard error; OR, odds ratio.

Although one study evaluated pesticide concentrations,¹¹⁾ the results were significant only for the reaction time test. Nevertheless, other tests had limited significance due to including a small size sample size. Regarding the meta-analysis results, exposure to pesticides was not associated with stunting. Therefore, the meta-analysis proved the potential bias in the retrieved papers that supports the null hypothesis.

Due to concerns about specific pesticides, namely organophosphate and carbamate, the retrieved studies agreed that neither is associated with stunting. The study by Jaacks et al. verified the presence of specific pesticide metabolites and found insufficient evidence of an association between exposure and stunting. We argue that the detection method used in that study lacked sensitivity (only a 60% detection rate for all metabolite pesticides). Stunting is an independent variable of neurodevelopment delay^{36,37)} that is not correlated with exposure status, even with higher exposure to pesticides and higher socioeconomic status and maternal education.³²⁾ The study of Orozco et al.³¹⁾ corroborated that, in the areas with the most extensive pesticide use, the average HAZ score was in the normal range. However, this study included a limited sample size.

Permethrin usage for malaria eradication is not correlated with stunting. A cohort study by Arinaitwe et al.¹⁸⁾ of res-

pondents treated with ITN evaluated the association between stunting and malaria. Beyond the strength of the research method, stunting versus non-stunting was correlated with the incidence of malaria. However, whether ITN was associated with stunting was doubtful due to a lack of information about insecticide dose and detection method. Thus, the correlation of ITN with stunting requires further investigation. Similarly, the study of Friedman et al.³³⁾ examined the impact of ITN on malnutrition, including stunting. However, the study was also prone to the causation of ITN coverage with stunting. Here, pyrethroid used for malaria eradication contributed to stunting. However, actual data to prove stunting is limited and unable to prove the causal relationship. It would be useful to determine whether malnutrition predicts infection in children or an infected child is prone to malnutrition and delayed development.³⁷⁻³⁹⁾ Further assessments are warranted to observe the causal

relationship.

This study has several limitations. The retrieved studies do not holistically verify multifactorial stunting, including protein intake, parental knowledge of stunting, educational background, and economic status. Most studies did not perform individual assessments of pesticide exposure; thus, it is difficult to verify the dose-response relationship. Community exposure by questionnaire or self-reporting is susceptible to selection bias. A robust study of pesticide exposure is limited, and the remaining included studies were Prone to bias (Table 3), especially in terms of allocation concealment and blinding in RCT studies as well as low-quality reporting on the GRADE summary (Table 4). Additionally, 75% heterogeneity and $P<0.000$ existed in the meta-regression analysis due to the age covariate (Table 5). Therefore, the effect of pesticide exposure on the incidence of stunting is uncertain and requires a completed assessment in

Table 3. Risk of bias (RoB) of included studies

NOS	Selection			Comparability				Exposure/outcome				Total
Jaacks et al. ²⁹⁾	***			**				***				8
Arinaitwe et al. ¹⁸⁾	***			**				***				8
Kartini et al. ¹²⁾	****			**				**				8
Paudel et al. ³⁰⁾	****			**				**				8
AHRQ	Q1	Q2	Q3	Q4	Q5	Q6	Q7	Q8	Q9	Q10	Q11	Total
Grandjean et al. ¹¹⁾	+	?	-	-	+	+	+	+	+	+	+	8
Friedman et al. ³³⁾	+	+	+	+	+	+	?	+	-	+	-	8
Amoako Johnson ³⁴⁾	+	-	?	+	?	+	?	+	NA	+	-	5
Orozco et al. ³¹⁾	+	?	+	-	?	+	-	+	-	-	-	3
Handal et al. ³²⁾	+	+	?	-	+	+	-	+	+	+	-	7
Cochrane RoB	Selection		Selection	Performance		Detection		Attrition		Reporting		Others
Friedman et al. ²⁸⁾	+		?	-		+		+		+		+
Ter Kuile et al. ¹⁷⁾	?		-	?		?		+		+		+

NOS, Newcastle-Ottawa Scale; AHRQ, Agency for Healthcare Research and Quality.
 , 2 criteria for RoB; *, 3 criteria for RoB; +, yes; -, no; ?, unclear.

Table 4. GRADE summary for certainty

No. of studies (No. of events/total participants)	Study limitation	Consistency	Directness	Precision	Publication bias	Quality
Pesticide exposure on stunting						
4 (197/468)	Not serious	Serious inconsistency (-1)	Direct	Imprecision (-1)	Unlikely	Low
Organophosphate exposure on stunting						
3 (397/707)	Not serious	Serious inconsistency (-1)	Direct	Imprecision (-1)	Unlikely	Low
Pyrethroid exposure on stunting						
8 (1,820/4,957)	Limitation	Serious inconsistency (-1)	Direct	Imprecision	Unlikely	Very low

GRADE, Grading of Recommendations Assessment, Development, and Evaluation.

Table 5. Meta-regression result of potential effect modifiers in meta-analysis

Variable	Estimate	SE	Z value	P value	95% CI
Intercept	0.966	0.310	3.115	0.002	0.358-1.574
Age	-0.025	0.005	-5.379	<0.001	-0.034 to -0.016
Sex (girl)	3.148	2.224	1.416	0.157	-1.210 to 7.507
Age*sex	-0.027	0.022	-1.238	0.216	-0.071 to 0.016

SE, standard error; CI, confidence interval.
 Regression model $0.966 + -0.025 \text{ age} + 3.148 \text{ sex (female)} + -0.027 \text{ age} \times \text{sex}$.

a forthcoming study. A future study must focus on multiple pesticide exposures, particularly of the organochlorine class, to evaluate the dose-response relationship of individual exposure with a large sample and using a robust method. IGF-1 testing should also be assessed to obtain accurate parameters of exposed groups.

In conclusion, beyond the agreement about the mechanism by which pesticides, especially organochlorine, influence stunting by inhibiting IGF-1, this study's findings do not suggest that pesticide exposure is correlated with stunting. Several studies denoted that pesticide exposure was correlated with stunting, but each was limited by a small sample size, lack of exposure assessment method, and poor design. Most published studies examined organochlorine pesticide-induced stunting in animal models; thus, the mechanisms by which organophosphate, pyrethroid, and carbamate affect stunting remain to be elucidated. The different toxicological properties of these pesticides will likely result in different mechanisms of action on growth hormone. Although determining the relationship between exposure and stunting is challenging, pesticides have negative effects on human health. This study provides evidence needed for a recommendation to control pesticide use in farming. Protecting vulnerable populations, including children in agricultural areas, is still urgently needed to prevent harmful events in forthcoming years.

Footnotes

Supplementary material: Supplementary Table 1 and Fig. 1 can be found via <https://doi.org/10.3345/cep.2023.01522>.

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